

Original Research

Effect of Using Chemical Fertilizers on Natural Radioactivity Levels in Agricultural Soil in the Iraqi Kurdistan Region

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Abstract

In this study, the effect of the use of chemical fertilizers on natural radioactivity levels in agricultural soil was investigated. For this purpose, virgin and agricultural (fertilized) soils were collected from Erbil city, the capital of the Iraqi Kurdistan Region. Gamma-ray spectrometry was applied using a high-purity germanium (HPGe) detector to determine the activity concentrations of ²²⁶Ra, ²³²Th, and ⁴⁰K in soil samples. The results showed that the range of activity concentrations for ²²⁶Ra, ²³²Th, and ⁴⁰K in virgin soils are (10.6-16.2) Bq/kg, (8.8-10.7) Bq/kg, and (241.8-340.9) Bq/kg, respectively, while the range of activity concentrations for ²²⁶Ra, ²³²Th, and ⁴⁰K in agricultural soils are (11.9-18.2) Bq/kg, (8.8-12.4) Bq/kg, and (247.7-338.3) Bq/kg, respectively. The obtained values of activity concentration show that the application of chemical fertilizers elevated the natural radioactivity level of the soil. The radiological hazard parameters were also calculated in both the virgin and agricultural soil samples and compared with the international dose safety limit in the soil.

Keywords: radioactivity, soil, effective dose, fertilizer, hazard

Introduction

In recent years, studies of natural radioactive nuclides in the environment and their role in the organisms of our planet have attracted rising attention [1, 2]. Also, natural radioactive materials have become of great interest in International Atomic Energy Agency (IAEA) publications and reports issued by the European Council Directive (EU) [2].

Natural radionuclides existing in the soil, water, and rocks are not distributed homogeneously throughout the world, but their concentrations depend mainly on geological and geographical conditions [3]. The natural radioactivity levels in soils have gained research interest because humans are exposed to natural radioactivity depending on the concentrations of these radioactive nuclides in each region of the world [4].

Information on radioactive nuclide distribution in soils is essential to the affected population and to control health risks [5]. The sources of radioactive nuclides are natural and artificial. The naturally occurring radioactive nuclides originate from ²³⁸U, ²³⁵U, and ²³²Th decay chains, and from ⁴⁰K, while artificial

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sources come from various applications of radionuclides in industries, nuclear weapons tests, and consumer products [7].

The concentration of radioactive nuclides in soil increases by adsorption with soil and its rainfall on soil, while the concentration decreases via the process of leakage and also dilutes when soil water content and organic matter increases the behaviour of radioactive nuclides in soil regarding site characteristics, amount and rate of rainfall and soil drainage [8].

The natural radioactive nuclides in soil vary from one global region to another. In soil, one of the sources of radioactive materials other than those of natural origin is mainly due to the wide use of chemical fertilizers for agricultural purposes [9]. Chemical fertilizers are used by farmers in the agricultural fields to reach high agricultural productivity. The compounds commercially of chemical fertilizers are named NPK (nitrogen (N), phosphorus (P) and potassium (K)). The ^{238}U and ^{232}Th enrichment in chemical fertilizers are linked to the high concentration in phosphate rock and the complicated chemical process in fertilizer production. The primary source of phosphate fertilizer is phosphate rocks of sedimentary origin, which contain relatively high concentrations of ^{238}U and ^{232}Th and its decay products [10, 11]. In NPK fertilizers, the potassium component increases the natural radioactivity, because of the presence of radioactive ^{40}K , whose natural abundance in potassium (K) is 0.0118% [2].

Moreover, fertilizer materials such as phosphates (which contain ^{238}U and ^{232}Th) and potassium, which are used in plant nitrification processes, are considered to be important sources of soil contamination and become a source of radioactivity [11]. This phenomenon may result in potential radiological risks owing to external exposure during a resident time in the farms and internal exposure through ingestion of food grown on fertilizer soils. Using phosphate fertilizers over a period of many years could eventually increase the radium and uranium

content of the soil and consequently increase radiation dose, which would result in the corresponding increase of the dose and cause diseases for the human body [12].

This study aims to assess the activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in virgin and fertilized soils collected from fields in Erbil city, where a variety of chemical fertilizers are being used by farmers to enhance crop yield. In addition, the use of chemical fertilizers affects radioactivity levels in the agricultural soil. Comparison between the results obtained for virgin and fertilized soils are presented.

Materials and Methods

Experimental Work

The present study was carried out in agricultural fields in Erbil (the capital of Iraq's Kurdistan Region). Four different locations (which consist of four greenhouses and 11 agricultural fields) in Erbil were selected for collecting the soil samples to determine the activity concentrations of naturally occurring radioactive nuclides in virgin and fertilized soils. The geographical position of sampling locations is shown in Fig. 1.

The selection of the fields was such that it contained soils that had received different amounts of fertilizers for different intervals of time. The plant fertilizers urea (N), diammonium phosphate (DAP), and urea ammonium phosphate (NP) were used by farmers in Ankawa, Binaslawa, and Bardarash locations to enhance the crop yield. Also, the plant fertilizers urea ammonium phosphate (NP), nitro phosphate potash mixed (NPK), calcium nitrate (CaN), and potassium sulphate (K) were used by farmers in the Yaremja (greenhouse and uncovered cultivate) location. In our previous work [8], the activity concentration of radioactive nuclides in

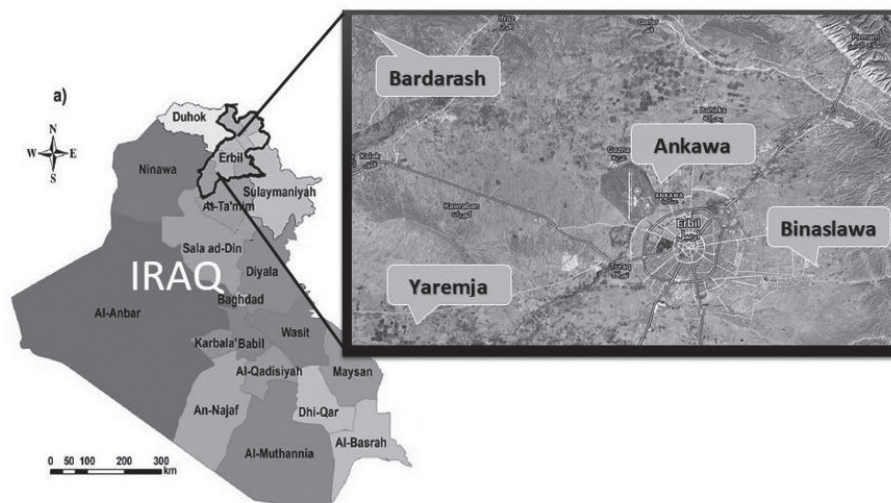


Fig. 1. Map of the studied area and sampling sites.

Table 1. Chemical composition of plant fertilizer and activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in Plant fertilizers used by the farmers in greenhouse and agricultural fields under study [11].

Name of plant fertilizer	Chemical analysis of fertilizers			Activity concentration (Bq/Kg)		
	N %	P ₂ O ₅ %	K ₂ O	^{226}Ra	^{232}Th	^{40}K
Potassium Sulfate (K)	0	0	52	1.0±0.2	0.2±0.02	12000±600
Urea (N)	46	0	0	0.2±0.04	0.2±0.03	0.4±0.03
Diammonium Phosphate (DAP)	18	46	0	8.0±1.0	1.5±0.1	14±1
Calcium Nitrate (16.5 %Ca + N)	13.5	0	0	2.0±0.2	0.2±0.05	1±0.2
Urea Ammonium Phosphate (NP) (1)	20	20	0	0.6±0.1	4.0±0.2	53±2
Urea Ammonium Phosphate (NP) (2)	18	44	0	0.4±0.13	10.0±1.0	4±0.5
N.P.K. Complex (1)	12	6	18	134.0±6.0	1.0±0.2	4000±200
N.P.K. Complex (2)	15	8	26	1.0±0.04	0.2±0.04	5000±250
N.P.K. Complex (3)	20	20	20	0.5±0.2	1.0±0.3	2400±120

these fertilizers was measured and analyzed, as shown in Table 1.

The soil samples were collected using a core method with a core of diameter 15 cm and a depth of 20 cm. The depth of the soil is considered to be especially important when the distribution of radionuclides is inhomogeneous along the soil depth [13]. After removing stones and inorganic materials, the soil samples were dried in an electric oven at about 120°C, before crushing and sieving through a 2mm mesh sieve. One kilogram of each sample was packed in a Marinelli beaker (one-litre size) for gamma spectrometry system and closed for six weeks to establish secular equilibrium between the radium contents of the samples and its daughter radionuclides [14, 15].

Data Acquisition

In this study, the gamma spectrometry system with HPGe setup and a multichannel analyzer was used for counting the gamma-rays emitted from soil samples. The high-purity germanium detector is p-type of vertical closed-end coaxial, manufactured by PGT (Princeton Gamma-Tec PGT Company-USA) [16]. The system was calibrated for energy using three standard sources: ^{60}Co , ^{137}Cs , and ^{226}Ra . The efficiency calibration was achieved using the same three standard sources. The detector was situated in a lead well with a thickness of 10cm to shield the measuring station versus background radioactivity.

The samples were placed over the detector for at least 10 hours. The spectra were evaluated using a Thermo Scientific System 8000 multi-channel analyzer and the Quantum Gold 2001 computer software program from PGT Company-USA. In order to determine the background radiation in the cavity around the detector, an empty beaker was counted, loaded and tested for a measurement time of 10 hours. The background spectra were used to correct the net peak area of gamma rays of measured isotopes.

After the measurement and the subtraction of the background, the naturally occurring (radioactive nuclides

^{226}Ra , ^{232}Th , and ^{40}K are considered in the spectrum analysis for the measured gamma-rays. The ^{226}Ra activity concentration was measured as the weighted average of the activity determined using the gamma-ray lines 351.9 keV (35.8%) gamma-rays from ^{214}Pb decay, 609.3 keV (44.8%), 1120 keV (14.8%), and 1764.5 keV (15.36%) gamma-rays from ^{214}Bi decay. The activity concentration of ^{232}Th , however, was measured as the weighted average of the activity determined using the gamma-ray lines 238.6 keV (43%) from ^{212}Pb decay, 583 keV (84.5%) and 2614.5 keV (99.16%) from ^{208}Tl decay, and 911.2 keV (26.6%) from ^{228}Ac decay. Furthermore, the activity concentration of ^{40}K was directly determined using the gamma-ray line at 1460.8 keV (10.7%).

Calculations

Activity Concentrations

Activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in soil samples were calculated using the following formula [2, 17]:

$$\text{Activity concentration} = (\text{Net Count} / (\epsilon \times I_{\gamma} \times T \times M)) \pm (\text{SD} / (\epsilon \times I_{\gamma} \times T \times M)) \quad (1)$$

...where ϵ is the absolute gamma peak efficiency for the detector at a particular gamma-ray energy, I_{γ} is the decay intensity of the specific energy peak (including the decay branching ratio information), T is the counting time for the measurement in seconds, M is the mass of the sample in kg, and SD is the standard deviation of the net count rate per second.

The percentage increase in activity concentrations of isotopes in agriculture soils was calculated using the following formula:

$$\text{Percentage increase in activity concentration (PIAC)} = [(AA - AV) / AV] \times 100\% \quad (2)$$

...where A_A and A_V are activity concentrations of the isotopes in agriculture and virgin soils, respectively.

Radium Equivalent Activity (Ra_{eq})

The distribution of natural radioactive nuclides in soils is not uniform. Therefore, the total exposure to radiation from ^{226}Ra , ^{232}Th , and ^{40}K nuclides was expressed by radium equivalent activity (Ra_{eq}) in (Bq/kg). The radium equivalent activity in soil samples was calculated using the following formula [18, 19]:

$$Ra_{eq} = (A_{Ra}) + (A_{Th} \times 1.43) + (A_K \times 0.077) \quad (3)$$

...where A_{Ra} , A_{Th} , and A_K represent the activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in a unit (Bq/kg), respectively. The safe value of Ra_{eq} for any naturally occurring radioactive materials is reported by OECD 1979 to be less than 370 Bq/kg [20, 21].

Absorbed Gamma Dose Rates (D_R)

The absorbed gamma dose rates in the air at 1 m above the ground surface for the uniform distribution of naturally occurring radionuclides were calculated based on guidelines provided by UNSCEAR [22].

$$D_R \text{ (nGy/h)(NORM)} = (A_{Ra} \times 0.462) + (A_{Th} \times 0.604) + (A_K \times 0.0417) \quad (4)$$

Annual Effective Dose

The annual effective dose (AED) was calculated from the absorbed gamma dose rates (D_R) by using the dose conversion factor of 0.7 Sv/Gy with an outdoor occupancy factor of 0.2 and 0.8 for indoor [22], and determined using the following equations [5, 23]:

$$\text{AED } (\mu\text{Sv/year}) = D_R \text{ (nGy/h)} \times T \times F \quad (5)$$

...where D_R is the calculated dose rate (in nGy/h) and T is the occupancy time. F is the conversion factor – a value is published by UNSCEAR as 0.7 Sv/Gy for environmental exposure to gamma rays of moderate energy [22, 24]. The outdoor occupancy factor T is about (20% of 8760 h/year). The outdoor annual effective dose equivalent is given by the following equation:

$$\text{AED}_{\text{outdoor}} \text{ } (\mu\text{Sv/year}) = D_R \text{ (nGy/h)} \times (0.2 \times 8760 \text{ h/year}) \times 0.7 \text{ (Sv/Gy)} \quad (6)$$

External Hazard Index (H_{ex})

The external hazard index (H_{ex}) is an estimation of the hazard of the natural gamma radiation. H_{ex} was determined by assuming that 370 Bq/kg of ^{226}Ra , 259 Bq/kg of ^{232}Th , and 4810 Bq/kg of ^{40}K are produced

at the same gamma dose rate. In order to evaluate this index, we can use the following relationship [5, 25]:

$$H_{ex} = (A_{Ra} / 370) + (A_{Th} / 259) + (A_K / 4810) \leq 1 \quad (7)$$

The external hazard index for ^{226}Ra , ^{232}Th , and ^{40}K nuclides is reported to be less than 1 mSv/year, which relates to a radium equivalent activity of 370 Bq/kg (OECD-1979) [20].

Excess Lifetime Cancer Risk (ELCR)

Excess lifetime cancer risk measures the additional cancer risk induced by exposure to ionizing radiations. Based on the calculated values of AED, ELCR is calculated using equation [26, 27]:

$$\text{ELCR}_{\text{outdoor}} = \text{AED}_{\text{outdoor}} \text{ } (\mu\text{Sv/year}) \times \text{DL} \times \text{RF} \quad (8)$$

...where AED is the annual effective dose, DL is the average duration of life (70 years), and Rf is the risk factor given as 0.05 by ICRP [28].

Results and Discussion

Activity Concentrations of Radioactive Nuclides

To assess the effect of the use of chemical fertilizers on the radioactivity levels in agricultural soils from Erbil governorate farms, the activity concentrations of natural radioactive nuclides ^{226}Ra , ^{232}Th , and ^{40}K in virgin and agricultural soils samples were measured using a gamma-ray spectrometer equipped with a high-purity germanium (HPGe) detector (Table 2).

The results showed that the mean values and the ranges of activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in virgin soils are 14.3 ± 2.2 (10.6-16.2) Bq/kg, 10.0 ± 0.8 (8.8-10.7) Bq/kg, and 290.2 ± 36.6 (241.8-340.9) Bq/kg, respectively. While the mean values and the ranges of activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in agricultural soils are 16.0 ± 1.8 (11.9-18.2) Bq/kg, 10.3 ± 1.1 (8.8-12.4) Bq/kg, and 294.6 ± 31.5 (247.7-338.3) Bq/kg, respectively. The variations in the activity concentrations of naturally occurring radioactivity nuclides in the soils in various locations in the studied area depend on the geological and geographical conditions in the area, as well as on the extent of fertilizer applied to agricultural lands because agricultural soils in the studied area had received different amounts of fertilizers for different intervals of time.

Fig. 2 shows the activity concentrations of ^{226}Ra and ^{232}Th in virgin and agricultural soils samples. Clearly, the fertilized soil samples are slightly more radioactive

Table 2. Activity concentrations of ²²⁶Ra and ²³²Th in virgin and agricultural soil samples.

Location	Code of Samples	Location	Activity concentration (Bq/Kg)		
			²²⁶ Ra	²³² Th	⁴⁰ K
Yaremja 1 (Green house)	SV 1	Virgin Soil	14.8±0.7	8.8±0.5	241.8±12.8
	SA 2	Soil from Agricultural field 1	16.7±0.8	8.8±0.5	262.6±15.6
	SA 3	Soil from Agricultural field 2	15.1±0.6	9.2±0.5	274.3±16.8
	SA 4	Soil from Agricultural field 3	15.0±0.7	8.8±0.4	247.7±13.1
Yaremja 2 (uncovered cultivate)	SV 5	Virgin Soil	14.8±0.7	10.6±0.6	300.8±15.9
	SA 6	Soil from Agricultural field 4	17.9±0.8	10.9± 0.6	311.2±16.4
	SA 7	Soil from Agricultural field 5	16.3±0.8	10.7±0.3	336.5±17.7
	SA 8	Soil from Agricultural field 6	18.2±0.8	12.4±0.7	335.8±18.0
	SA 9	Soil from Agricultural field 7	15.0±0.7	10.9±0.6	306.2±16.1
Ankawa	SV 10	Virgin Soil	16.2±0.6	10.7±0.5	295.0±15.4
	SA 11	Soil from Agricultural field 8	17.0±0.8	10.9±0.6	286.0±15.1
Binaslawa	SV 12	Virgin Soil	15.3±0.5	10.2±0.4	272.5±14.2
	SA 13	Soil from Agricultural field 9	15.3±0.6	10.4±0.5	272.1±14.3
	SA 14	Soil from Agricultural field 10	16.6±0.6	10.3±0.5	264.2±13.8
Bardarash	SV 15	Virgin Soil	10.6±0.3	9.8±0.5	340.9±18.3
	SA 16	Soil for Agricultural field 11	11.9±0.6	10±0.5	338.3±17.6

in comparison with the soil samples in which no fertilizer was used (virgin soil) for Ankawa, Bardarash, and Bnaslawa locations, while for Yaremja location (in both greenhouses and uncovered cultivated fields) the variation is more clear because in this location different types of chemical fertilizers such as potassium sulfate (K), urea ammonium super phosphate (NP), and nitro phosphate potash mixed (NPK) fertilizer were used for different intervals of time. These chemical fertilizers, especially phosphate fertilizers, are enriched by ²²⁶Ra

and ²³²Th, and its decay products are linked to the high concentration in phosphate rock and the chemical process involved in fertilizer production, as shown in Table 1.

Fig. 3 shows the activity concentration of ⁴⁰K in virgin and agricultural soil samples. In this figure, it is clear that the activity concentration of ⁴⁰K in the fertilized soil samples collected from agricultural fields in Yaremja (in both greenhouses and uncovered cultivated fields) is higher than the soil samples in which

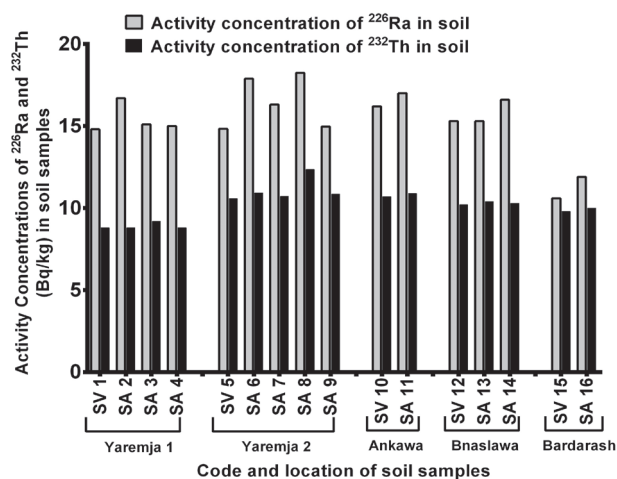


Fig. 2. Activity concentrations of ²²⁶Ra and ²³²Th in virgin (SV) and agricultural (SA) soil samples.

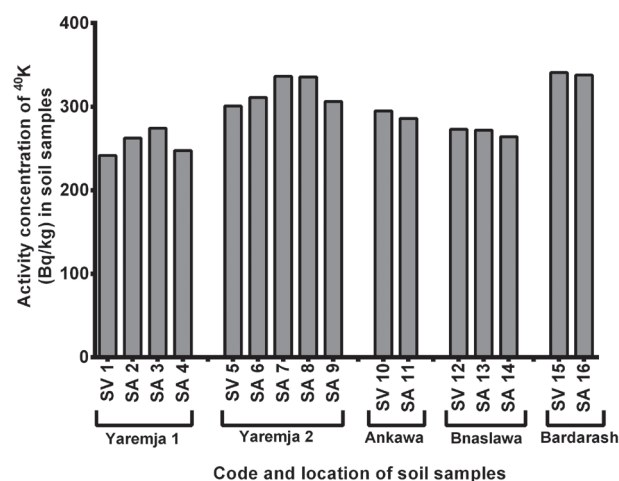


Fig. 3. Activity concentration of ⁴⁰K in virgin (SV) and agricultural (SA) soil samples.

Table 3. Percentage ratio of increase in activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in agricultural soil samples due to the use of chemical fertilizers in this field.

Location	Code of Samples	Location	Increasing in the activity concentration %		
			^{226}Ra	^{232}Th	^{40}K
Yaremja 1 (Green house)	SV 1	Virgin Soil			
	SA 2	Soil from Agricultural field 1	12.5	0.2	8.6
	SA 3	Soil from Agricultural field 2	1.6	4.5	13.4
	SA 4	Soil from Agricultural field 3	1.2	0.2	2.4
Yaremja 2 (uncovered cultivate)	SV 5	Virgin Soil			
	SA 6	Soil from Agricultural field 4	20.6	3.2	3.5
	SA 7	Soil from Agricultural field 5	10.1	1.4	11.9
	SA 8	Soil from Agricultural field 6	23.0	16.9	11.6
	SA 9	Soil from Agricultural field 7	1.4	2.6	1.8
Ankawa	SV 10	Virgin Soil			
	SA 11	Soil from Agricultural field 8	5.3	1.9	0
Binaslawaw	SV 12	Virgin Soil			
	SA 13	Soil from Agricultural field 9	0.5	1.9	0
	SA 14	Soil from Agricultural field 10	8.9	1.1	0
Bardarash	SV 15	Virgin Soil			
	SA 16	Soil for Agricultural field 11	12.8	1.9	0

Table 4. Activity concentrations of naturally occurring radioactive nuclides in virgin and agriculture soils in the present work are compared with results reported from other countries throughout the world.

Country	Activity concentration of NORM in virgin soil (Bq/kg)			Activity concentration of NORM in agriculture soil (Bq/kg)			References
	^{226}Ra	^{232}Th	^{40}K	^{226}Ra	^{232}Th	^{40}K	
Brazil				3-21	0.4-17	31-124	[30]
Italy				24-231	20-70	242-1434	[31]
Greece				8-68	8-78	185-868	[32]
Algeria	47	33	329	53	50	311	[8]
Egypt				2-14	2-14	54-170	[33]
Pakistan				30-50	42-56	591-709	[34]
Malaysia	45-111	52-127	99-173	59-167	88-181	202-529	[5]
India	3-16	37-299	338-544	3-21	48-309	322-564	[29]
China	9-145	15-102	417-1263				[35]
KSA				10-19	9-28	542-773	[6]
Turkey	13-31	12-37	285-614				[36]
Turkey				60-98	55-81	450-737	[37]
Iran	20	23	613	29-38	26-34	630-816	[38]
Iraq	16-39	9-28	262-613				[39]
Iraq	11-16	9-11	242-341	12-18	9-12	248-338	Present work
Range in worldwide	16-110	11-64	140-850	16-110	11-64	140-850	[22]
Worldwide average	35	30	400	35	30	400	[22]

Table 5. Radium equivalent activity (Ra_{eq}), external absorbed gamma dose rates (D_R), outdoor annual effective dose (AED), external hazard index (H_{ex}), and excess lifetime cancer risk (ELCR) for virgin and agricultural soil.

Location	Code of Samples	Location	Ra_{eq} (Bq/kg)	D_R (nGy/h)	AED (μ Sv/year)	H_{ex} (mSv/year)	(ELCR) outdoor
Yaremja 1 (Green house)	SV 1	Virgin Soil	46.0±2.4	22.2±1.2	27.3±1.4	0.12±0.01	9.5 E-5
	SA 2	Soil from Agricultural field 1	49.5±2.7	24.0±1.3	29.4±1.6	0.13±0.01	1.0 E-4
	SA 3	Soil from Agricultural field 2	49.3±2.7	23.9±1.3	29.3±1.6	0.13±0.01	1.0 E-4
	SA 4	Soil from Agricultural field 3	46.7±2.3	22.6±1.1	27.7±1.4	0.12±0.01	9.7 E-5
Yaremja 2 (uncovered cultivate)	SV 5	Virgin Soil	53.1±2.7	25.8±1.3	31.6±1.6	0.14±0.01	1.1 E-4
	SA 6	Soil from Agricultural field 4	57.5±2.9	27.8±1.4	34.1±1.7	0.16±0.02	1.2 E-4
	SA 7	Soil from Agricultural field 5	57.6±2.6	28.0±1.3	34.4±1.6	0.16±0.02	1.2 E-4
	SA 8	Soil from Agricultural field 6	61.8±3.1	29.9±1.5	36.7±1.9	0.17±0.03	1.3 E-4
	SA 9	Soil from Agricultural field 7	54.1±2.8	26.2±1.3	32.2±1.7	0.15±0.02	1.1 E-4
Ankawa	SV 10	Virgin Soil	54.3±2.6	26.3±1.2	32.2±1.5	0.15±0.03	1.1 E-4
	SA 11	Soil from Agricultural field 8	54.6±2.8	26.4±1.3	32.3±1.6	0.15±0.01	1.1 E-4
Binaslawa	SV 12	Virgin Soil	50.9±2.1	24.6±1.0	30.2±1.3	0.14±0.01	1.1 E-4
	SA 13	Soil from Agricultural field 9	51.2±2.4	24.7±1.2	30.3±1.4	0.14±0.01	1.1 E-4
	SA 14	Soil from Agricultural field 10	51.8±2.4	24.9±1.2	30.6±1.4	0.14±0.02	1.1 E-4
Bardarash	SV 15	Virgin Soil	50.9±2.5	25.0±1.2	30.7±1.5	0.14±0.02	1.1 E-4
	SA 16	Soil for Agricultural field 11	52.3±2.7	25.7±1.3	31.5±1.6	0.14±0.03	1.1 E-4
Average			52.9±4.1	25.6±2.0	31.5±2.4	0.14±0.02	1.1 E-4

no fertilizer was used (virgin soil), because in these locations potassium content fertilizers such as potassium sulfate (K) and nitro phosphate potash mixed (NPK) fertilizer were used to enhance the crop yield, which are rich in potassium radioactive isotope ^{40}K . On the other hand, a comparison of the virgin and fertilized soils of the Ankawa, Binaslawa, and Bardarash locations reveals that there is no increase in the activity concentrations of ^{40}K in both types of soils because, as shown in Table 1, the fertilizers used in these areas were urea, urea ammonium phosphate (NP), and diammonium phosphate (DAP); there is no potassium content in these fertilizers.

Table 3 shows an increase in activity concentration percentage ratios of ^{226}Ra , ^{232}Th , and ^{40}K in agricultural soil samples due to the use of chemical fertilizers in these fields. The data shows that the use of chemical fertilizers in the studied area has elevated the activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K by factors of 0.5-23%, 0.2-16.9%, and 0-13.4%, respectively.

The obtained values of activity concentrations of naturally occurring radioactive nuclides in both virgin and agricultural soil samples were within the world average values of 35, 30, and 400 Bq/kg for ^{226}Ra , ^{232}Th , and ^{40}K , respectively [22]. Activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in soil samples from the area under study are compared with results reported from other countries throughout the world (Table 4).

Radiological Hazard Assessment

In order to estimate the radiological hazards, we calculated the radium equivalent activity (Ra_{eq}), external hazard index (H_{ex}), external absorbed gamma doses rate (D_R), outdoor annual effective dose (AED)_{outdoor}, and excess lifetime cancer risk (ELCR) for virgin and agricultural soil samples (Table 5).

The calculated radium equivalent activity (Ra_{eq}) in virgin soil samples varied from (46.0-54.3) Bq/kg, and in agricultural soil samples varied from (46.7-61.8) Bq/kg, with the average value (52.9±4.1 Bq/kg). Variations in the radium equivalent activity in different soil samples depended on the type and content of natural radioactive nuclide as well as the type of chemical fertilizer used in the studied area. In Fig. 4, it is clear that the radium equivalent activity in the virgin soil sample is lower than the soil samples collected from agricultural fields where fertilizers are used. Thus the use of chemical fertilizers in the fields to enhance crop yield promotes the radium equivalent activity, and therefore the exposure of the farmers working in these fields. The obtained values of radium equivalent activity in soil samples are within the recommended limits of 370 Bq/kg [20, 21, 29].

The absorbed gamma dose rates (D_R) were calculated to vary from (22.2-26.3) nGy/h in virgin soil samples, while in soil samples collected from agricultural fields

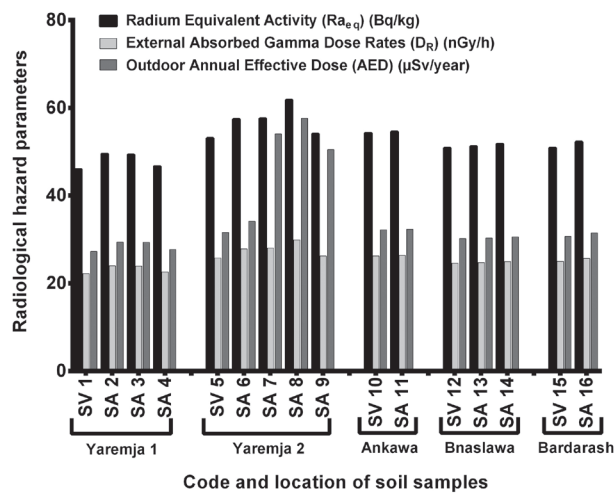


Fig. 4. Radium equivalent activity (Ra_{eq}), external absorbed gamma dose rates (D_R), and outdoor annual effective dose (AED) in virgin (SV) and agricultural (SA) soil samples under study.

vary from (22.6-29.9) nGy/h, with an average value of 25.6 ± 2.0 nG/h. The calculated average value of the absorbed gamma dose rates (D_R) was lower than the worldwide average value of 59 nGy/h as published by UNSCEAR [22].

The calculated outdoor annual effective dose (AED_{outdoor}) in virgin soil samples varied from (27.3-32.2) μSv/year, and in agricultural soil samples varied from (27.7-36.7) μSv/year, with the average value (31.5 ± 2.4 μSv/year). The annual effective dose values are within the worldwide average value of 70 μSv/year, as published by UNSCEAR [22]. As shown in Fig. 4, the calculated absorbed gamma dose rates and annual effective dose equivalent for the soil samples in which fertilizers were used are higher than the virgin soil samples.

The external hazard index ranges for virgin and agricultural soil samples were (0.12-0.15) mSv/year, and (0.12-0.17) mSv/year, respectively. The calculated external hazard index for the soil samples in which fertilizers were used are higher than the virgin soil samples only in Yaremja (in both fields greenhouses and uncovered cultivate), because in these fields different types of chemical fertilizers (such as potassium sulfate (K), urea ammonium super phosphate (NP), and nitro phosphate potash mixed (NPK) fertilizer) were used, while there is no change observed in values of external hazard index for both virgin and agricultural soil samples in Ankawa, Binaslaw, and Bardarash, where chemical fertilizers urea, diammonium phosphate, and urea ammonium phosphate were used in these fields to enhance the crop yield. The external hazard index for the soil samples in the studied area is lower than the safety limit of 1 mSv/year as recommended by the OECD [20, 29].

The calculated excess lifetime cancer risk (ELCR_{outdoor}) values for virgin and agricultural soil samples are (9.5×10^{-5} - 1.1×10^{-4}), and (9.7×10^{-5} - 1.3

$\times 10^{-4}$), respectively. From the comparison of the values of excess lifetime cancer risk for virgin and agricultural soil samples of the investigated region, there is no drastic change in the excess lifetime cancer risk in both types of soils.

The results of the present study indicate that the use of chemical fertilizers in agriculture fields to enhance crop yield increases the activity concentrations of natural radioactive nuclides and therefore the exposure of the farmers working in the fields, especially potassium sulfate (K), urea ammonium super phosphate (NP), and nitro phosphate potash mixed (NPK) fertilizers with high concentrations of ^{226}Ra , ^{232}Th , and ^{40}K radioisotopes.

Conclusions

The activity concentrations of naturally occurring radioactive nuclides in virgin and agricultural soil of the studied area were measured using gamma spectrometry. The obtained data revealed that the activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K in agricultural soil were higher than those in virgin by factors of (0.5-23.0)%, (0.2-16.9)%, and (0-13.4)%, respectively. The use of chemical fertilizers in the agricultural fields to enhance the crop yield causes an increase in the activity concentrations of natural radioactive nuclides, as well as the exposure of the farmers working in the agricultural fields, especially in those fields using potassium sulfate (K), urea ammonium super phosphate (NP), and nitro phosphate potash mixed (NPK) fertilizers with high concentrations of ^{226}Ra , ^{232}Th , and ^{40}K . The values of the activity concentrations of naturally occurring radionuclides in the virgin and the fertilized soil of the studied area were within the world average as proposed by UNSCEAR.

The calculated values of the radiological hazard parameters (radium equivalent activity, the external hazard index, external absorbed gamma dose rates, outdoor annual effective dose, and excess lifetime cancer risk) for virgin and agricultural soil samples were in general also increased and modified by the effect of fertilizer use. However, in both virgin and fertilized soils, the level of these parameters was below the international dose limit in the soil as proposed by UNSCEAR and OECD. The conclusion of this work is that these farms and their soils and subsequently their products pose no health hazard for the population in terms of radiation impact.

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