



Supplement of

Chemical composition, optical properties, and oxidative potential of waterand methanol-soluble organic compounds emitted from the combustion of biomass materials and coal

Tao Cao et al.

Correspondence to: Jianzhong Song (songjzh@gig.ac.cn)

The copyright of individual parts of the supplement might differ from the article licence.

22	CONTENTS:
----	------------------

- 23 1. Biomass and coal samples
- 24 2. Collection of smoke samples from BB and CC
- 25 3. Extraction and fractionation of BrC
- 4. Organic carbon/elemental carbon (OC/EC) and total organic carbon (TOC) analysis
- 27 5. UV-visible properties
- 28 6. Principal component analysis (PCA)
- 29 7. Quality control
- 8. Table S1. Region, excitation/emission wavelength maxima range and attribution of
- 31 chromophores in BrC emitted from BB and CC
- 32 9. Table S2. Results of DTT assay conducted on the WSOC, HULIS and MSOC of smoke
 33 samples
- 10. Figure S1. The normalized UV-vis spectra by organic carbon contents of WSOC, HULIS,
- 35 and MSOC fractions
- 11. Figure S2. EEM fluorescence counter maps of corresponding WSOC, HULIS, MSOC of
- BB and CC smoke samples, presented as specific intensity (a.u. L(mg C-1))
- 12. Figure S3. ¹H-NMR stacking diagram of corresponding WSOC, HULIS, MSOC of BB
- and CC smoke samples. The segment from 4.40 to 5.60 ppm was removed for NMR
- 40 spectra due to MeOH and H2O residues. The peaks were assigned to specific compounds
- 41 as follows: Levoglucosan (L), Phthlic acid (PA)
- 42
- 43

44

S1. Biomass and coal samples

In this study, six biomass materials and five types of coal were collected and burned to 45 investigate the optical and chemical properties of brown carbon (BrC) fractions emitted from 46 biomass burning (BB) and coal combustion (CC) smoke. The six biomass materials consisted 47 of three types of crop straw (wheat straw [WS], rice straw [RS], and corn straw [CS]) and 48 three types of wood (pine wood [PW], Chinese fir [CF], and white poplar [WP]). The three 49 crop straws were chosen because they were the main types of crop straw burned in China. 50 These crop straws are usually used as fuels for heating in the winter or cooking in rural areas 51 52 throughout the year, and are also occasionally burned in agricultural fields after the harvest season (Ke et al., 2019). The three wood materials are widespread in forests and are 53 commonly used as biomass fuels in some rural areas of China. The combustion of these crop 54 55 straws and woods has been reported to make a significant contribution to the atmospheric aerosol in China (Fan et al., 2018; Shen et al., 2013). Therefore, these biomass materials were 56 selected as representative biomass fuels for the study of BB-derived BrC. In this study, WS, 57 RS, and CS were collected in the rural area of Bengbu, Anhui Province, China, while PW, CF, 58 and WP were collected from a forest area in Lu'an, Anhui Province, China. Before the 59 experiment, the crop straws and wood materials were washed with water and air dried for 60 seven days. 61

In some developing countries, such as China, coal is still an important fuel in rural areas and also makes a large contribution to the levels of atmospheric pollution. In this study, five coals were chosen for the investigation of their BrC fractions from CC. They consisted of four bituminous coals (B-1, B-2, B-3, and B-4, with volatile fractions of 34%, 32%, 25%, and

19%, respectively) and one anthracitic coal (with a volatile fraction of 3.3%). These five 66 coals represented the major types of coal used for residential CC in China. After collection, 67 the coals were washed with water three times to remove dust and then air-dried. Then raw 68 coal was crushed, fully mixed, and made into coal briquettes. 69

70

71

S2. Collection of smoke samples from BB and CC

Samples of the smoke emitted from BB and CC were collected in a combustion and 72 sampling system that was introduced in our previous studies (Fan et al., 2018; Li et al., 2018). 73 74 The instrument was made of stainless steel and consisted of a combustion hood, clean air dilution, and injection ports, smoke pipe, mixing fan, mixing chamber, PM_{2.5} sampler 75 (JCH-120F Intelligent medium flow PM sampler, Juchuang Environmental Protection Group 76 77 Co., Ltd, Qindao, China), and exhaust port. The smoke samples emitted from BB and CC were then collected as follows: 78

(1) Biomass burning smoke samples. The biomass fuels were first prepared as small 79 80 pieces (length ~10 cm) and then placed on a combustion stove. After dropping 1 mL of alcohol on the biomass fuels they were ignited with an electronic gas lighter. The smoke 81 particles were diluted and transported into the mixing chamber. Finally, smoke particles were 82 collected on quartz fiber filters (Ø 90 mm: Whatman, Maidstone, UK) in a PM_{2.5} sampler at a 83 flow rate of 80 L/min. Five complete experiments were conducted for each biomass fuel and 84 five smoke PM_{2.5} filter samples were obtained. 85

(2) Coal combustion smoke samples. The smoke particles emitted from the CC samples 86 were also collected in the same combustion and sampling system. Sample collection was 87

conducted according to the method introduced by Li et al. (2018). Briefly, two anthracite 88 briquettes were ignited in a burning coal honeycomb briquette stove and were then moved 89 90 into the other coal stove. After the burning stage of coal was reached and smoke emissions were minimized, the coal stove was placed into the sampling system. Then, one honeycomb 91 sample was placed in the pre-burned coal stove. The resulting smoke was diluted and passed 92 into the mixing chamber. Finally, smoke samples were collected with the PM_{2.5} sampler at a 93 flow rate of 80 L/min. To obtain sufficient smoke sample for the comprehensive 94 characterization of the BrC fractions, each coal was burned at least for three cycles. All 95 96 quartz filters were baked for 6 h at 450 °C to remove any organics absorbed on the filters and then wrapped with baked aluminum foil. After sampling, the filter samples were re-wrapped 97 with baked aluminum foil and stored in a refrigerator $(-20 \text{ }^{\circ}\text{C})$ prior to analysis. 98

Field blank quartz filters were collected before each group of combustion experiments under conditions in which the fuels were not ignited. The field blank filters were used to correct the mass of smoke $PM_{2.5}$ and water-/methanol- soluble BrC, as well as the optical signals and DTT consumption by BrC. To prevent contamination of the following sample, the collection system was cleaned before each new combustion experiment.

104

105 S3. Extraction and fractionation of BrC

The BrC fractions (i.e., water-soluble organic compounds [WSOC], humic-like substances [HULIS], and methanol-soluble organic compounds [MSOC]) were obtained with solvent extraction and a solid-phase extraction (SPE) method, as indicated in our previous studies (Fan et al., 2018; Fan et al., 2016; Li et al., 2018). The filter samples were

ultrasonically extracted three times with 20 mL ultrapure water for 30 min. The extracts were 110 filtered through a 0.22 µm polytetrafluoroethylene (PTFE) syringe filter to obtain the WSOC 111 112 fraction. The HULIS fraction in WSOC was isolated with an SPE method (Chen and Bond, 2010; Zhang et al., 2013; Cheng et al., 2016; Cheng et al., 2017). Briefly, the pH of the 113 WSOC solution was acidified to 2 with HCl, and the solution introduced into a 114 pre-conditioned SPE cartridge (Oasis HLB, 200 mg, Waters, Milford, MA, USA). The most 115 hydrophilic species and inorganic salt ions or metal ions was removed by the cartridge, 116 whereas the relatively hydrophobic HULIS fraction was retained. Then the SPE column was 117 118 rinsed with pure water to remove inorganics and the retained organics were eluted with methanol. Finally, the HULIS solution was evaporated to dryness under a gentle nitrogen 119 stream. It is noted that HULIS is less polar components of the WSOC, which were also 120 121 usually refer to the relatively hydrophobic fraction of water-soluble organic carbon (Verma et al., 2012, Zheng et al., 2013, Katsumi et al., 2018). 122

The MSOC was obtained by a method developed by Cheng et al. (2016). Briefly, the filter samples were immersed in 20 mL methanol (Macklin, >99.9%, Shanghai, China) for 2 h and then filtered through a 0.22 μ m PTFE syringe filter (Jinteng, Tianjin, China). Static digestion without ultrasonic treatment can avoid the loss of particulate matter and facilitate the determination of dissolved organic matter content. Finally, the dried residual filters and untreated filters were analyzed to determine their carbon contents.

129

130 S4. Organic carbon/elemental carbon (OC/EC) and total organic carbon (TOC) analysis
131 The OC and EC in smoke filter samples were measured using an OC/EC analyzer (TOT,

Sunset Laboratory Inc., Portland, OR, USA). The analysis was conducted according to the 132 National Institute of Occupational Safety and Health (NIOSH) 870 method (Chow et al., 133 2001; Wu et al., 2016). The TOC content of WSOC and HULIS was determined by a 134 high-temperature catalytic oxidation instrument (VCPH analyzer, Shimadzu, Kyoto, Japan) 135 following the non-purgeable OC protocol. After the removal of inorganic carbon, the sample 136 was oxidized at high temperature (680 °C) and the peak area of CO₂ was determined by a 137 non-dispersive infrared detector. The content of the MSOC fraction was indirectly obtained 138 by subtracting the TC concentrations of the extracted filters from that of the untreated filters. 139 140 The experiments were all repeated three times and the concentrations reported here were corrected for their respective blank concentrations. 141

In this study, the "µgC" was used as weight unit that referring to the mass of carbon for
the OC, EC, WSOC, HULIS-C, and MSOC fractions.

144

145 **S5. UV-visible properties**

The UV-visible absorption spectra of the BrC fractions (i.e., WSOC, HULIS, and MSOC) were recorded between the wavelengths of 200 to 700 nm using a UV-2600 UV-vis spectrophotometer (Shimadzu, Japan). The sample solution was placed in a 1-cm quartz cuvette and analyzed at 1 nm intervals. Ultrapure water was used as a blank reference for the WSOC and HULIS solutions, while pure methanol was used for the MSOC fraction. The field blank sample solution was also used as the blank sample, and the interference from the instrument and operating blank was determined.

153 The absorption Ångström exponent (AAE) is a measure of the spectral dependence of

the light absorption of BrC solutions (Cheng et al., 2016), which was calculated by thefollowing equation:

156
$$A_{\lambda} = K \lambda^{-AAE}$$
(2)

where A_{λ} is the absorbance derived from the spectrophotometer at a given wavelength λ (330– 400 nm) and *K* is a constant.

The mass absorption efficiency at 365 nm (MAE₃₆₅) is an important parameter used to characterize the light absorbing ability of BrC. It was obtained using the following equation:

161
$$MAE_{\lambda} = \frac{A_{\lambda}}{c \cdot L} \times \ln(10)$$
(3)

where A_{λ} is the absorbance at λ nm, *c* is the carbon concentration of BrC in solution (µgC mL⁻¹), and *L* is the absorbing path length.

165 S6. Principal component analysis (PCA)

PCA is a widely used chemometric procedure that can transform the original variables to 166 the principal component by dimension reduction analysis. The principal component (two or 167 three) generally contain most of the original variable information (Popovicheva et al., 2020). 168 Thus, in this study, PCA was performed to find out the key factors that may affect the DTT 169 activities from a series of characteristic of BrC fraction (MAE₃₆₅, percentage content of four 170 fluorophores, and percentage content of R-H, H-C-C=, H-C-O, and Ar-H groups). The main 171 calculation is based on SPSS version 19 (IBM SPSS Statistics) (Shivani et al., 2019), and the 172 confidence interval is 95%. 173

174

175 S7. Quality control

The experimental blank and field blank were both analyzed in this study. The experimental blank was treated in the same method as it for smoke samples, which was used to assess the errors that may be introduced in the process of experimental operation and their repeatability. In this study, the average blank values of sextuple collections of WSOC, HULIS and MSOC were $0.28 \pm 0.07 \ \mu gC/cm^2$, $0.12 \pm 0.03 \ \mu gC/cm^2$ and $0.38 \pm 0.09 \ \mu gC/cm^2$, respectively. The repeatability of analysis procedure was obtained based on one blank filter sample, which were 3.1%, 2.2% and 4.5% for WSOC, HULIS and MSOC, respectively.

The field blank filter (ambient and ignition coal (for coal combustion only)) were collected follow the procedure for sampling smoke $PM_{2.5}$ samples, but without ignited fuel samples. The field blank filters were treated as the method for smoke samples. The average values of WSOC, HULIS and MSOC were $1.8\pm0.2 \ \mu gC/cm^2$, $0.7\pm0.1 \ \mu gC/cm^2$ and $5.3\pm0.9 \ \mu gC/cm^2$, respectively. They were much less than the values of that in smoke particle.

In the present study, all the BrC results were blank-corrected by subtracting an average field blank value for each sample. The data were present as a mean \pm standard deviation based on triplicate analysis of filter sample for each combustion experiment.

191

Table S1. Region, excitation/emission wavelength maxima range and attribution of chromophores in BrC

emitted from BB and CC

	Region	$\lambda_{ex} \max(nm)$	$\lambda_{em} \max(nm)$	Fluorescent compounds	References
	Ι	220-250	290-320	protein-like amino acid	(Cui et al., 2016; Coble, 1996)
	II	220-250	320-380	protein-like UV region	(Mostofa et al., 2011; Mounier et al., 2010)
	III	220-250	380-460	fulvic-like	(Chen et al., 2003; Santos et al., 2012)
	IV	250-380	280-380	tryptophan-like/microbial	(Santos et al., 2012; Cui
	V	250-380	380-460	byproduct humic-like organic	et al., 2016) (Chen et al., 2003; Qin et al., 2018)
195 196					ct al., 2018)
197 198					
199 200					
201	Table S2. Results of DTT assay conducted on the WSOC, HULIS and MSOC of smoke samples				

		Calculated by PM mass(pmol/min/µg)		
	Samples	WSOC	HULIS	MSOC
Biomass	WS	4.5±3.8	3.2±2.8	85±12
burning	RS	6.1±0.5	5.5±0.6	84±5.6
	CS	7.4±1.4	3.0±0.7	69±11
	PW	5.9±3.3	3.1±0.3	9.1±1.5
	CF	5.5±2.3	3.2±0.9	14±6.9
	WP	5.6±2.8	2.6±1.0	11±7.6
Coal	B-1	1.6±0.2	1.1±0.1	7.7±0.8
combustion	B-2	2.1±0.2	1.5±0.1	11±3.2
	B-3	0.5±0.1	0.5±0.1	3.2±1.7
	B-4	1.9±0.5	0.9±0.2	3.1±1.0
	AN	0.7±0.2	0.4±0.1	6.7±2.9

^a error bars represent standard deviation based on quadruplicate test

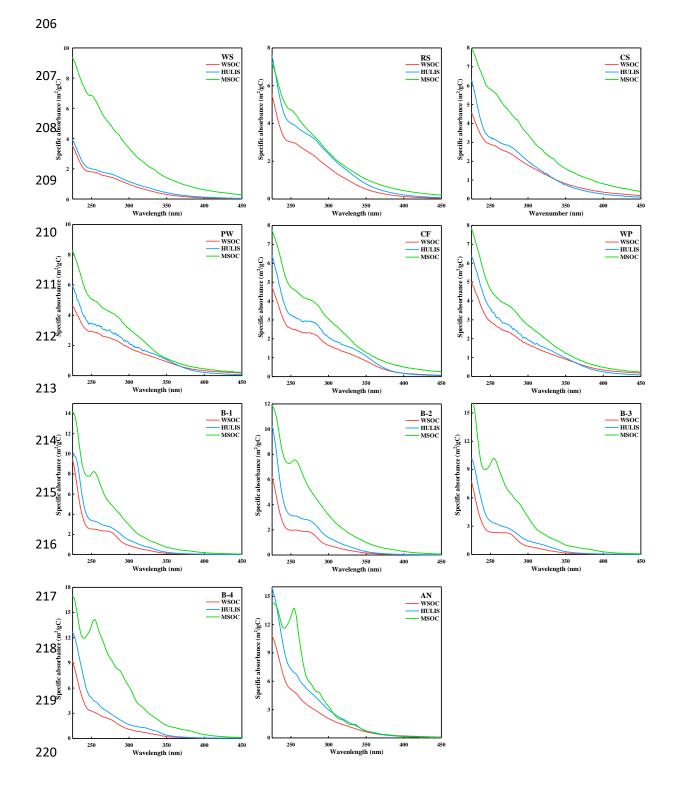
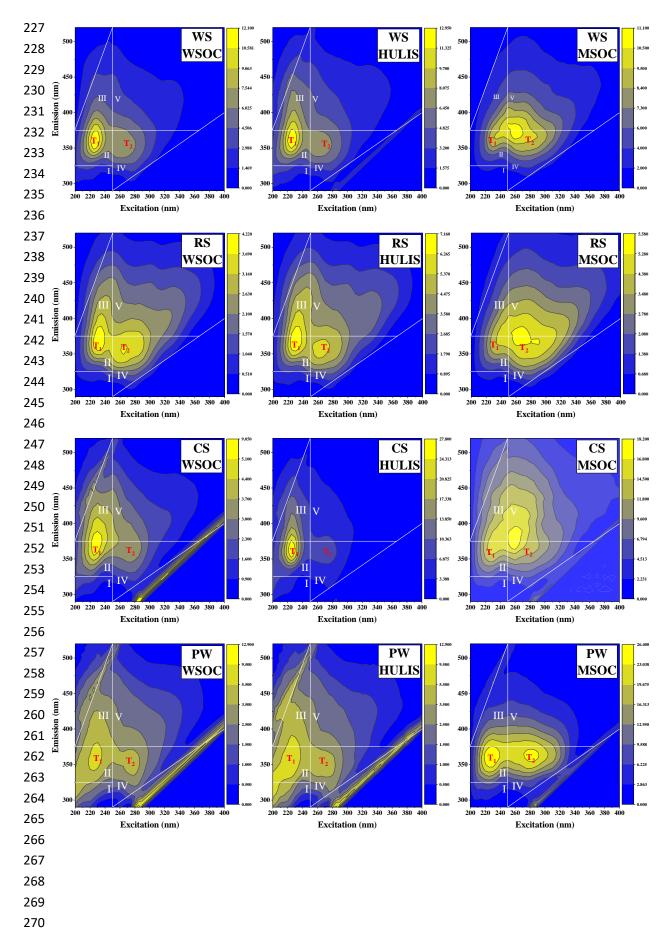
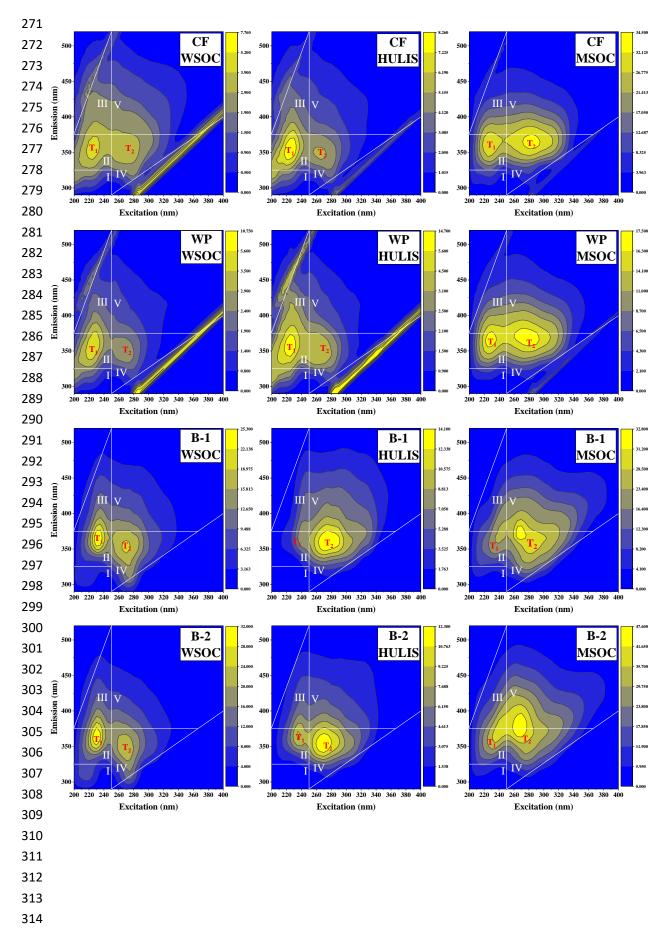
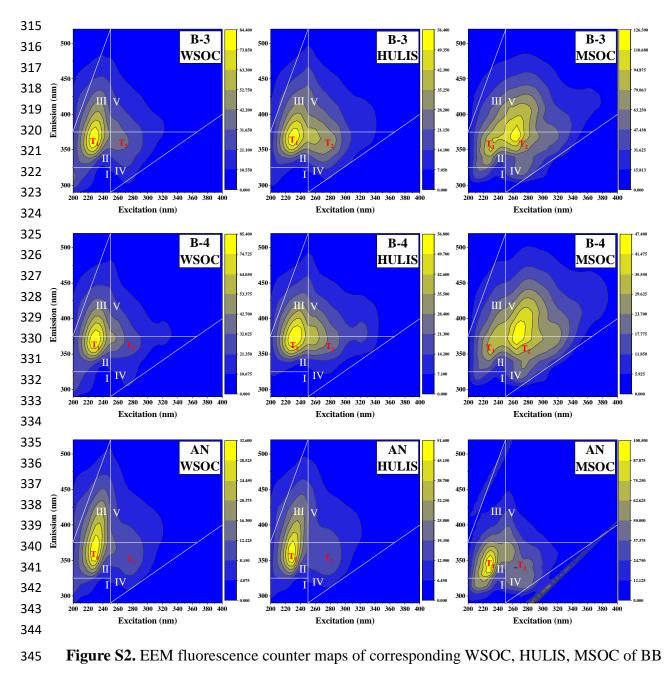


Figure S1. The normalized UV-vis spectra by organic carbon contents of WSOC, HULIS,

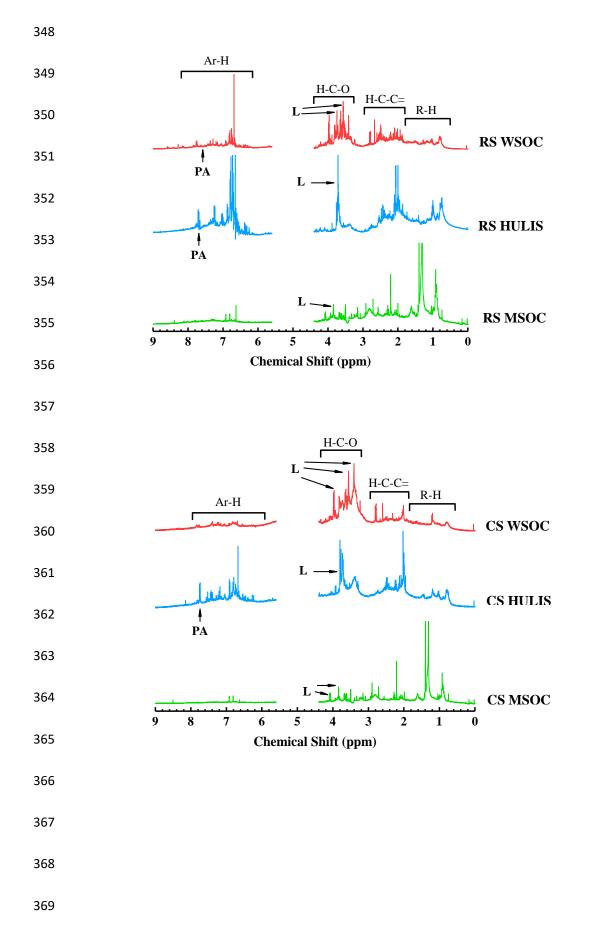
- and MSOC fractions
- 223
- 224
- 225
- 226

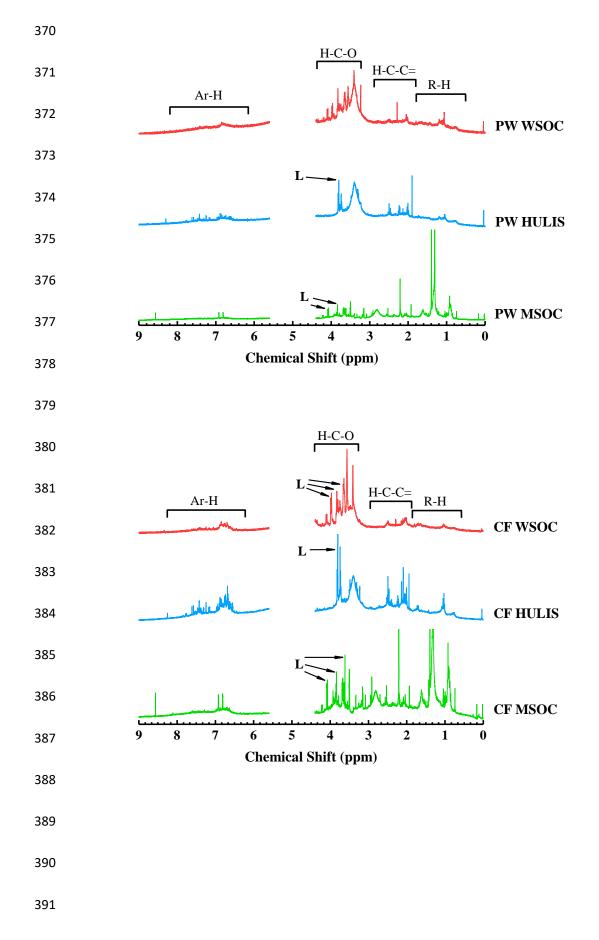


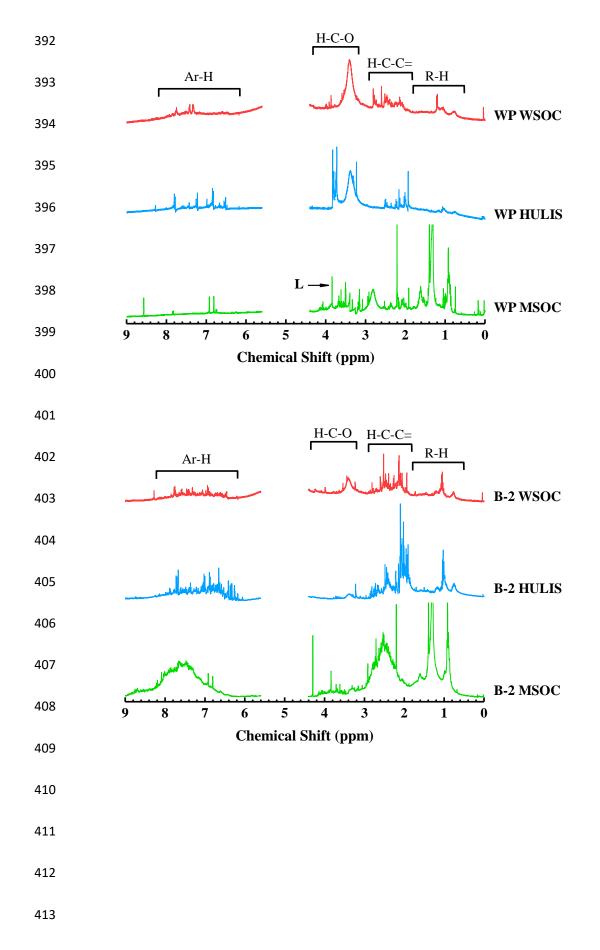


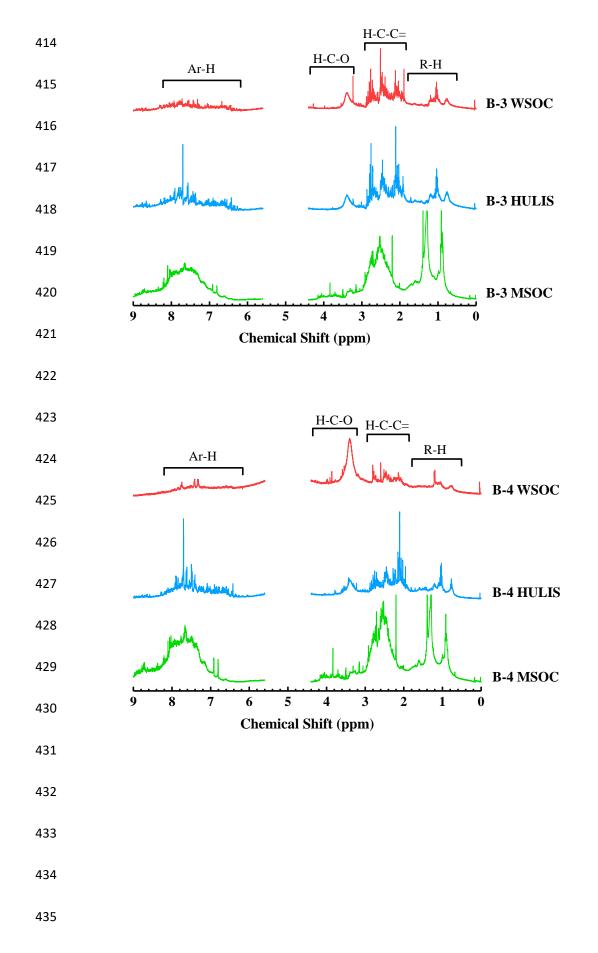


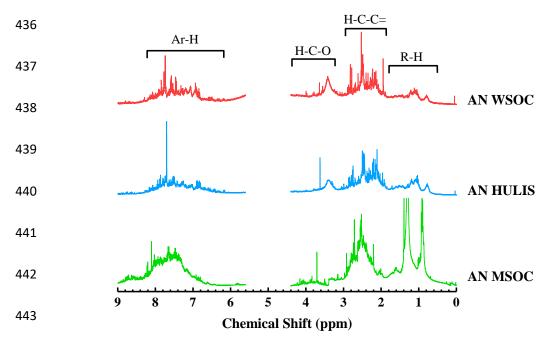
and CC smoke samples, presented as specific intensity (a.u. $L(mg C^{-1})$)











444

Figure S3. ¹H-NMR stacking diagram of corresponding WSOC, HULIS, MSOC of BB and
CC smoke samples. The segment from 4.40 to 5.60 ppm was removed for NMR spectra due
to MeOH and H₂O residues. The peaks were assigned to specific compounds as follows:
Levoglucosan (L), Phthlic acid (PA).

450 **References**

451	Chen, W., Westerhoff, P., Leenheer, J. A., and Booksh, K.: Fluorescence excitation - Emission
452	matrix regional integration to quantify spectra for dissolved organic matter,
453	Environmental science & technology, 37, 5701-5710, 10.1021/es034354c, 2003.
454	Chen, Y., and Bond, T. C.: Light absorption by organic carbon from wood combustion,
455	Atmospheric Chemistry and Physics, 10, 1773-1787, DOI 10.5194/acp-10-1773-2010,
456	2010.
457	Cheng, Y., He, K. B., Du, Z. Y., Engling, G., Liu, J. M., Ma, Y. L., Zheng, M., and Weber, R.
458	J.: The characteristics of brown carbon aerosol during winter in Beijing, Atmospheric
459	Environment, 127, 355-364, 10.1016/j.atmosenv.2015.12.035, 2016.
460	Cheng, Y., He, K. B., Engling, G., Weber, R., Liu, J. M., Du, Z. Y., and Dong, S. P.: Brown
461	and black carbon in Beijing aerosol: Implications for the effects of brown coating on
462	light absorption by black carbon, The Science of the total environment, 599-600,
463	1047-1055, 10.1016/j.scitotenv.2017.05.061, 2017.
464	Chow, J. C., Watson, J. G., Crow, D., Lowenthal, D. H., and Merrifield, T.: Comparison of
465	IMPROVE and NIOSH carbon measurements, Aerosol Science and Technology, 34,
466	23-34, 10.1080/027868201300081923, 2001.
467	Coble, P. G.: Characterization of marine and terrestrial DOM in seawater using excitation
468	emission matrix spectroscopy, Marine Chemistry, 51, 325-346,
469	10.1016/0304-4203(95)00062-3, 1996.
470	Cui, X., Zhou, D., Fan, W., Huo, M., Crittenden, J. C., Yu, Z., Ju, P., and Wang, Y.: The
471	effectiveness of coagulation for water reclamation from a wastewater treatment plant

- that has a long hydraulic and sludge retention times: A case study, Chemosphere, 157,
 224-231, 10.1016/j.chemosphere.2016.05.009, 2016.
- 474 Fan, X., Li, M., Cao, T., Cheng, C., Li, F., Xie, Y., Wei, S., Song, J., and Peng, P. a.: Optical
- properties and oxidative potential of water-and alkaline-soluble brown carbon in smoke
- 476 particles emitted from laboratory simulated biomass burning, Atmospheric Environment,
 477 194, 48-57, 10.1016/j.atmosenv.2018.09.025, 2018.
- Fan, X. J., Wei, S. Y., Zhu, M. B., Song, J. Z., and Peng, P. A.: Comprehensive
 characterization of humic-like substances in smoke PM2.5 emitted from the combustion
 of biomass materials and fossil fuels, Atmospheric Chemistry and Physics, 16,
- 481 13321-13340, 10.5194/acp-16-13321-2016, 2016.
- Katsumi, N., Miyake, S., Okochi, H., Minami, Y., Kobayashi, H., Kato, S., Wada, R.,
 Takeuchi, M., Toda, K., and Miura, K.: Humic-like substances global levels and
 extraction methods in aerosols, Environmental Chemistry Letters, 17, 1023-1029,
 10.1007/s10311-018-00820-6, 2018.
- 486 Ke, H., Gong, S., He, J., Zhou, C., Zhang, L., and Zhou, Y.: Spatial and temporal distribution
- of open bio-mass burning in China from 2013 to 2017, Atmospheric Environment, 210,
- 488 156-165, 10.1016/j.atmosenv.2019.04.039, 2019.
- Li, M., Fan, X., Zhu, M., Zou, C., Song, J., Wei, S., Jia, W., and Peng, P.: Abundances and
 light absorption properties of brown carbon emitted from residential coal combustion in
 China, Environmental science & technology, 10.1021/acs.est.8b05630, 2018.
- 492 Mostofa, K. M. G., Wu, F. C., Liu, C. Q., Vione, D., Yoshioka, T., Sakugawa, H., and Tanoue,
- 493 E.: Photochemical, microbial and metal complexation behavior of fluorescent dissolved

494	organic matter in the aquatic environments, Geochem. J., 45, 235-254, 2011.
495	Mounier, S., Zhao, H., Garnier, C., and Redon, R.: Copper complexing properties of
496	dissolved organic matter: PARAFAC treatment of fluorescence quenching,
497	Biogeochemistry, 106, 107-116, 10.1007/s10533-010-9486-6, 2010.
498	Popovicheva, O., Ivanov, A., and Vojtisek, M.: Functional Factors of Biomass Burning
499	Contribution to Spring Aerosol Composition in a Megacity: Combined FTIR-PCA
500	Analyses, Atmosphere, 11, 319, 10.3390/atmos11040319, 2020.
501	Qin, J., Zhang, L., Zhou, X., Duan, J., Mu, S., Xiao, K., Hu, J., and Tan, J.: Fluorescence
502	fingerprinting properties for exploring water-soluble organic compounds in PM 2.5 in an
503	industrial city of northwest China, Atmospheric Environment, 184, 203-211,
504	10.1016/j.atmosenv.2018.04.049, 2018.
505	Santos, P. S., Santos, E. B., and Duarte, A. C.: First spectroscopic study on the structural
506	features of dissolved organic matter isolated from rainwater in different seasons, The
507	Science of the total environment, 426, 172-179, 10.1016/j.scitotenv.2012.03.023, 2012.
508	Shen, G., Chen, Y., Wei, S., Fu, X., Zhu, Y., and Tao, S.: Mass absorption efficiency of
509	elemental carbon for source samples from residential biomass and coal combustions,
510	Atmospheric Environment, 79, 79-84, 10.1016/j.atmosenv.2013.05.082, 2013.
511	Shivani, Gadi, R., Sharma, S. K., and Mandal, T. K.: Seasonal variation, source
512	apportionment and source attributed health risk of fine carbonaceous aerosols over
513	National Capital Region, India, Chemosphere, 237, 124500,

10.1016/j.chemosphere.2019.124500, 2019. 514

Verma, V., Rico-Martinez, R., Kotra, N., King, L., Liu, J., Snell, T. W., and Weber, R. J.: 515

516 Contribution of water-soluble and insoluble components and their 517 hydrophobic/hydrophilic subfractions to the reactive oxygen species-generating 518 potential of fine ambient aerosols, Environmental science & technology, 46, 519 11384-11392, 10.1021/es302484r, 2012.

- 520 Wu, C., Huang, X. H. H., Ng, W. M., Griffith, S. M., and Yu, J. Z.: Inter-comparison of 521 NIOSH and IMPROVE protocols for OC and EC determination:implications for
- 522 inter-protocol data conversion, Atmospheric Measurement Techniques, 9, 4547-4560,
- 523 10.5194/amt-9-4547-2016, 2016.
- Zhang, X., Lin, Y. H., Surratt, J. D., and Weber, R. J.: Sources, composition and absorption
 Angstrom exponent of light-absorbing organic components in aerosol extracts from the
 Los Angeles Basin, Environmental science & technology, 47, 3685-3693,
 10.1021/es305047b, 2013.
- Zheng, G., He, K., Duan, F., Cheng, Y., and Ma, Y.: Measurement of humic-like substances
 in aerosols: a review, Environmental pollution, 181, 301-314,
 10.1016/j.envpol.2013.05.055, 2013.

531